Rational nitrogen fertilization in intensive cropping systems

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Abstract

The objective of a rational N fertilization program is to account for the sources and fate of N while estimating crop N needs. Efficiency of N use will vary with cropping systems and N sources. Management technologies that affect N use efficiency include the amount of N applied, timing and placement of N fertilizer, and use of inhibitors. One of the main problems in making a fertilizer N recommendation is to account for the contribution of N mineralization to plant available N. Most laboratory procedures do not account for the environmental factors that affect N mineralization and only estimate the size of the mineralizable N pool. However, changes in soil moisture and temperature can dramatically affect the amount and rate of release of mineralized N. Field and modeling techniques are two possible techniques to estimate N mineralization. Field techniques can be divided into soil and plant approaches. Soil incubations in the field provide a quantitative approach while soil nitrate tests during the growing season provide a qualitative approach to estimating N mineralization. The plant is the ultimate integrator of N mineralization. Plant N uptake by an unfertilized crop can provide a quantitative approach with certain precautions. This approach may be costly, labor intensive, and site specific. Crop N uptake during the growing season can be estimated by measuring the tissue N content or using a chlorophyll meter. The chlorophyll meter measures the greenness of the plant and has been shown to be positively correlated to plant N status. Modeling may provide another option by including the factors that affect the rate of N mineralization from a known pool. The two most important variables include soil moisture and temperature. Realistic yield expectations and accounting for existing and projected amounts of available N can improve the accuracy of N recommendations.

Introduction

Production of food and its protein content are key components to sustaining the world population. In many cropping systems, N frequently limits crop yield and protein levels and additional N inputs are required to optimize productivity and profitability. Nitrogen availability to crops is regulated by a biologically dynamic soil N cycle; thus, it is subject to the environmental factors that regulate the activity of microorganisms. In general, a portion of the N applied in excess of crop requirements ultimately will be converted to NO₃, which may be leached to groundwater. Thus, it is important to develop rational practices for N fertilizer to ensure sufficient but judicious use of N for economic and environmental reasons. Previous N recommendations were based on inexpensive fertilizer, a

lack of appreciation for spatial variability, and minimal environmental concerns. In some environmentally sensitive areas, future N recommendations need to be as site specific as possible.

Accurate recommendations for N fertilizer require knowledge of plant N requirements, external N sources, and N losses from the soil system. These components are represented by the equation:

$$N_f = N_c - (N_{sources} - N_{losses}) \tag{1}$$

where:

 $N_f = N$ fertilizer

 $N_c = N$ needed by the crop

 $N_{sources} = N \text{ sources}$

 $N_{losses} = N losses$

This apparently simple equation is very complex. The amount of N needed by the crop (Nc) is a function of dry matter yield and N concentration of the crop. The estimate of crop N requirement influences N recommendations more than any other factor. Overestimating the yield goal will result in a N recommendation that exceeds the actual crop N requirement; thus, establishing a realistic yield goal is important for reducing the environmental impact of N use (Jackson et al., 1987). Power and Broadbent (1989) suggested using a yield goal that is 5 to 10% greater than average yield of the previous 5 to 7 years. It is unrealistic to expect precise estimates because of yearly variations in growing season weather. If the yield includes the entire plant, then the N concentration of that yield will determine the crop's N needs. In the case where only a portion of the crop is harvested (i.e. grain), an estimate of the N requirement for the entire crop should be used.

Exogenous N inputs include fertilizer, manure, and biologically fixed N. The contribution of biologically fixed N becomes more important in tropical systems, where free-living and associative N_2 fixation increases. Associative N2 fixation with tropical grasses, including sugar cane, has been reported to be as much as 100 kg N ha-1 (Neyra and Dobereiner, 1977; Dobereiner, 1978). In rice production, the N contribution of Azolla has been reported to be equivalent to 30 kg urea-N ha-1 (Watanabe, 1987). Inputs from irrigation and precipitation also must be considered. Nitrogen inputs from precipitation normally are small, but localized sources can make a significant contribution. Nitrogen inputs from irrigation also are normally small, but when NO3 contaminates the groundwater, significant amounts of N can be added through irrigation. Schepers and Mosier (1991) reported that $> 100 \text{ kg N ha}^{-1} \text{ can be added dur-}$ ing one growing season in the Platte River valley of Nebraska in the US, where the NO3-N concentration can exceed 30 mg N L⁻¹.

Internal inputs include residual soil inorganic N and N mineralized from crop residues, soil organic matter, and manure that decompose during the growing season. In perennial crops, root N that will be translocated to the shoots also should be considered.

Nitrogen losses include leaching of NO_3^- , gaseous N losses from soil through denitrification, volatilization of N from fertilizers and vegetation, and erosion. The quantity of N lost through these pathways greatly depends on the local conditions of soil, crop, and climate

Our simple equation now takes the form:

$$N_f = N_c - [(N_m + N_{in} + N_r + N_{fix}) - (N_l + N_d + N_v + N_c) - N_{im}]$$
(2)

where,

 $N_m = N$ mineralization from soil organic matter

 N_{in} = residual inorganic soil N

 $N_r = N$ from residue

 N_{fix} = biologically fixed N

 $N_l = N$ loss from leaching

 $N_d = N$ loss from denitrification

 $N_{v} = N$ loss from volatilization

 $N_c = N$ loss from erosion

 N_{im} = temporary N loss from immobilization

The ultimate goal of a N management program is to accurately supply the crop requirement, while minimizing N losses through leaching, denitrification, volatilization, and erosion. Minimizing N losses and maximizing crop recovery of applied N will enhance N use efficiency and reduce environmental risk inherent with the use of external N inputs.

Efficiency of N use

Implicitly included in a N fertilizer recommendation is the efficiency of N use by the crop. Each of the N sources identified in Equation 2 will have an associated efficiency that is a function of losses. Management systems that dramatically change N use efficiency require that the N recommendation model be recalibrated. Nitrogen use efficiency is a function of climate; soil properties; crop and soil management; and management of the N applied (time, form, and placement). Although we often have very little control of climate, changes in precipitation amounts and timing will alter N transformations and subsequent availability. Myers (1988) reported N fertilizer efficiencies ranging from 12 to 74% for arable crops in the tropics. Poor recovery could be due to leaching and denitrification in wet climates and poor N uptake in dry climates. Crop and soil management practices are important considerations in management of N inputs, e.g. the effect tillage systems have on fertilizer transformations and microbial activity. For example, Phillips et al. (1980) reported that no-tillage corn production required more N fertilizer compared to plowed systems. This conclusion is based

Crop Growth - Maize

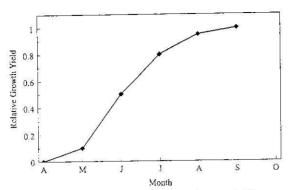


Fig. 1. Typical growth of maize in the central US.

on the premise that fertilizer N is used less efficiently under a no-tillage system. In reality, no-tillage systems generally exhibit higher denitrification (Rice and Smith, 1982) and greater potential for immobilization (Rice and Smith, 1984) and leaching (Thomas et al., 1973). Crop recovery of applied N sometimes can be enhanced by adopting new technologies that improve the synchrony between crop needs and N supply.

Nitrogen management technologies have been developed to reduce N contact with soil microorganisms, N leaching, and N volatilization. Reducing N contact with soil microorganisms can reduce denitrification and immobilization. Nitrification inhibitors also can be used with N sources to temporarily reduce microbial activity responsible for the oxidation of NH₄⁺. Nitrogen management technologies incompass application timing and placement of fertilizers and manures, use of inhibitors and N amount.

Timing

Maximum crop recovery and use of applied N require that N available be synchronous with plant growth and N demand (Fig. 1). Nitrogen synchrony often is not practical or possible. Supplying N after the crop reaches full canopy without damage to the crop can be difficult unless it is supplied by fertigation, a high-clearance vehicle, or aerial application. The timing of N availability must consider N mineralization as well as external N inputs. McGill and Myers (1987) discussed this relationship using a moisture-temperature index as an indicator of microbial activity and N mineralization. For example, the semi-arid tropical climate

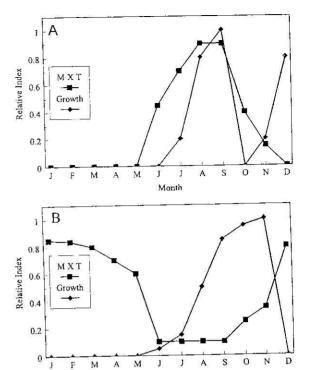


Fig. 2. Moisture-temperature index (■) for N mineralization and relative growth (♦) for (A) a semi-arid, tropical site with kharif and rubi at Hyderabad, India and (B) a subtropical site with wheat at Darling Downs, Australia (McGill and Myers, 1987).

Month

of Hyberabad, India, creates conditions conducive for synchrony between N mineralization and kharif and rabi crops (Fig. 2A). An example of a nonsynchronous relationship between N mineralization and crop N demand occurs in sub-tropical Queensland Australia with spring wheat (Fig. 2B). To improve the recovery of fertilizer N, some producers will split their N fertilizer application or delay the application into the growing season. This practice can successfully increase N fertilizer use efficiency and potentially reduce N fertilizer inputs. Studies have shown a 10 to 30% decrease in the amount of N fertilizer needed by the crop from this practice (Wells, 1984).

Placement

Adoption of N placement technologies can reduce N losses and increase plant N uptake. Placement of N becomes more important as reduced tillage systems are adopted. In the United States of America, common placement methods for maize include either sur-

Table 1. Fertilizer N recovered by corn as affected by tillage (1985-1986 average)

| N Placement | Time Applied | Tillage System (%) | |
|----------------|-----------------|--------------------|---------|
| | | Plowed | No-till |
| Broadcast | At planting | 46 | 36 |
| | Delayed | 56 | 44 |
| Injected | At planting | 59 | 52 |
| | Delayed | 56 | 56 |

Pierce and Rice (pers. commun.).

face or subsurface banding of the N fertilizer (Lamond et al., 1991). Placement of fertilizer N is even more critical with the use of urea in rice production. Volatilization losses are extremely high (up to 96%) unless deep placement of urea is used (Watanabe et al., 1981; Zhao-liang, 1981). Subsurface placement of manure N can reduce N volatilization losses of surfaceapplied manure. Often, timing and placement interact positively (Table 1). These data show greater fertilizer N recovery by maize when the N application was delayed to the sixth leaf stage or subsurface banded at planting. In this example, no benefit resulted from subsurface banding with a delayed application. Other studies have shown the benefit of banded fertilizer on recovery of applied N (Murphy et al., 1978; Mengel, 1982).

Inhibitors

Specific inhibitors of biological N processes have been developed and can be used to increase N use efficiency and decrease leaching losses. The most common and effective products inhibit nitrification and urea hydrolysis. Nitrification inhibitors temporarily inhibit the nitrification process when crop N uptake is low and denitrification and leaching potentials are high. Several compounds have been reported to inhibit nitrification, but only three have been developed commercially: nitrapyrin, N serve, [2-chloro-6-(trichloromethyl)pyridine]; etridi-[5-thoxy-3-(trichloromethyl)-1,2,4-Dwell, azol. thiadiazaole]; and dicyandiamide, DCD, (Peterson and Frye, 1989). The extent of nitrification inhibition depends upon the amount and time of N application, soil temperature, and precipitation patterns (Frye, 1981). Generally, conditions conducive for leaching and denitrification will enhance the effectiveness of nitrification inhibitors. A more complete discussion

Table 2. Grain yield response of maize to urease inhibitor

| Treatment | Irrigated | Dryland | |
|---------------------------------|------------------------|---------|--|
| | (Mg ha ⁻¹) | | |
| Urea | 7.02a | 3.16a | |
| Urea + NBPT | 8.28b | 4.30b | |
| NH ₄ NO ₃ | 8.28b | 3.85b | |

Lamond et al. (1994).

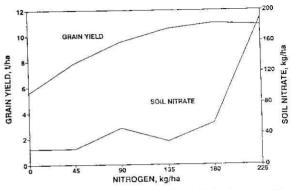


Fig. 3. Nitrogen response curve for maize. Soil nitrate concentrations in the surface 3.2 m. N rate for economic optimum yield was about 180 kg N ha^{-1} .

of nitrification inhibitors is presented by Sahrawat (1994).

Urease inhibitors reduce urease activity, resulting in delayed hydrolysis of urea or urea-based fertilizers that will increase transport of surface-applied urea into the soil. Movement of urea into the soil can significantly reduce volatilization losses of applied urea. This can translate to greater grain yields (Table 2) (Lamond et al., 1994). A more complete discussion of urease inhibitors is presented by Bremner (1994).

Amounts of N applied

Identifying the precise amount of external N applied to a crop is extremely important in maximizing crop recovery of N and minimizing environmental risk associated with N use. A typical N response curve is illustrated in Figure 3. The largest incremental response to applied N occurs at the lower amounts of N application. More importantly, the final 10% yield response to applied N uses approximately 35% of the N fertilizer required for maximum profit (Schlegel and Havlin, 1994). These data also show that increasing the amount

of N applied beyond the economic optimum amount dramatically increases profile N content and NO₃⁻ leaching potential.

One of the main problems in developing an accurate N recommendation is to quantify the contribution of N mineralization, because it is an important source of plant N. Accurately assessing N mineralization is difficult and has been the subject of research for almost 100 years. Universal procedures to quantify N mineralization still have not been developed. Early research focused on developing a laboratory test to determine the N mineralization potential of soils. Laboratory procedures can be divided into chemical and biological indices. These are reviewed and discussed elsewhere (Keeney, 1982; Meisinger, 1984; Rice and Havlin, 1994; Stanford, 1982). Generally, the laboratory procedures have not been adopted widely, because they do not predict N mineralization under field conditions. These laboratory procedures generally have been unsuccessful, because they fail to consider the dynamics of the factors that regulate the rate of N mineralization. These factors are substrate quality; moisture; substrate accessibility (clay content, tillage; wetting and drying cycles, freeze-thaw cycles); temperature; and pH. A complete discussion of these factors is given by Rice and Havlin (1994).

The nature and complexity of these regulatory factors complicate assessment of N mineralization in the laboratory. Field techniques and modeling are two other possible choices. Field techniques include soil sampling and incubations over time and plant sampling. Field techniques integrate many of the regulatory factors noted above, as well as crop and soil management factors that affect N mineralization. A disadvantage of the field techniques is that they are often site specific. Field techniques can be divided into quantitative or qualitative approaches and have been discussed in greater detail by Schepers and Meisinger (1994).

Assessment of nitrogen mineralization

Soil incubations

A quantitative approach to field N mineralization is the use of fallow soil. This approach avoids the disturbance that occurs with laboratory techniques. The lack of disturbance can emulate conditions that occur in the field with differing soil management systems, such as reduced tillage and residue management. Other physical conditions also are preserved. The soil in conservation tillage systems, often is more dense, cooler, and wetter than plowed soil. To eliminate large leaching and denitrification losses, the fallow technique can include a cover (Rice et al., 1987). Rice et al. (1987) reported differences between tillage systems and soil drainage classes under the covered fallow soils that were not shown by laboratory techniques. One potential disadvantage of covering the soil is that it would not undergo the same wetting and drying cycles that occur under normal field conditions. A more detailed discussion of this technique is presented by Schepers and Meisinger (1994).

Soil nitrate tests

Soil nitrate tests provide a relative index of in situ N mineralization. Specific soil NO₃⁻ tests have many variations that are dependent on climatic conditions, crop, and soil type. In the humid temperate regions of the US, the pre-sidedress nitrate test (PSNT) for maize is being evaluated. This test measures the amount of NO₃ in the top 30 cm of the soil profile when the corn is at the 6-leaf stage or approximately 30 cm tall (Bock et al., 1992). The PSNT was proposed first by Magdoff et al. (1984) to identify situations where fertilizer would not be needed to attain optimum maize yields. The PSNT integrates residual NO₃ from the previous season, N losses, and spring N mineralization until the time sampled. The intent of this test is to allow N mineralization to proceed as long as possible before the accumulated NO3 is measured and yet allow enough time to apply sidedress N before the crop grows too tall to prevent access to the field. The advantage is that, in the eastern two thirds of the US, the PSNT can separate sites that have sufficient N (> 20-25 mg NO₃-N kg⁻¹) from those that require additional fertilizer N inputs for maize production. The PSNT does not quantitatively measure N mineralization. For the PSNT to be useful, a standard sampling time and rapid analysis are important so that sidedress fertilizer N can be applied in a timely manner. Presently, the PSNT has been evaluated only for maize and needs to be evaluated for other crops. Neteson et al. (1989) reported that soil profile inorganic N currently is being used in the Netherlands to adjust N fertilizer recommendations for several crops. Including soil profile inorganic N was either equally effective or decreased N fertilizer requirement when compared to using a fixed N rate.

Plant N status

The plant is the ultimate integrator of environmental variables controlling the N transformations within the soil N cycle and, ultimately, the N available to it. Two approaches will be discussed: a qualitative approach involving chlorophyll meters and a quantitative approach using N uptake by an unfertilized crop. Crop N uptake provides an estimate of N availability during the growing season. The N status of the crop can be quantified by determining the tissue N content or using a chlorophyll meter. The chlorophyll meter measures the chlorophyll content or greenness of the plant that is correlated positively with crop N status. The chlorophyll meter is ideal for real time analysis of the crop N status and indicates N stress. The principle is discussed further by Schepers et al. (1992a,b) who had success comparing chlorophyll meter readings of an area requiring N to an area where N is nonlimiting because of application of adequate N fertilizer. Some of the disadvantages discussed by Schepers and Meisinger (1994) are that the data generated are site and plant (variety) specific. This makes calibration of the meters difficult unless a sufficiency index is calculated.

Plant N uptake

Nitrogen uptake by an unfertilized crop represents the best method for quantifying net N mineralization. The advantages of this approach are that it integrates field temperature, moisture, and aeration conditions that greatly influence N mineralization potential and can be adapted easily to include many crop and soil management practices. Nitrogen uptake in unfertilized plots also includes the normal rooting depth of the crop being studied and integrates spatial and temporal influences on N mineralization. Accurate measures of crop N uptake must include total biomass N and not just grain yield. One problem with estimating total N mineralization is accounting for root N content. A constant shoot:root N ratio cannot always be assumed and must be validated. Other disadvantages of this technique are costs, labor requirements, and site specificity. Information derived from one site may not be transferable to another site, unless some other nonsite-specific parameters can be correlated to N mineralization or can be used to predict it.

Models

Modeling techniques are other options that could directly incorporate the factors affecting microbial activity and N mineralization, as well as the other factors affecting N availability and losses. Current attempts to model N mineralization for field situations include the rate factor (k) and the size of the mineralizable pool (No). Most studies suggest use of first order kinetics (Broadbent, 1986; Juma et al., 1984; Stanford and Smith, 1972). A few studies have suggested a two-pool model to predict N mineralization (Molina et al., 1980; Deans et al., 1986). For use under field conditions, the rate then is adjusted for moisture and temperature (Campbell et al., 1981; Myers et al., 1982). However, Campbell et al. (1988) reported that their model underestimated N mineralization during the growing season. They attributed the underestimate to flushes of N mineralization during rewetting of dry soils. For temperature, Honeycutt et al. (1988) suggested using accumulated heat units to predict N mineralization; however, the heat unit concept is unable to predict N mineralization under dry soil conditions (Doel et al., 1990). Such models need to include climatic parameters and soil properties. Neeteson et al. (1989) have successfully used a N fertilizer recommendation model for potatoes in the Netherlands.

Fertilizer N recommendation

Fertilizer N recommendations need to be calibrated across broad areas, even though site-specific information is used to develop the appropriate algorithms. To calibrate N recommendations, knowledge of the crop response to N fertilizer is required. Critical variables for determining the N rate need to be identified and measured. Important site variables include texture, pH, and organic matter content. Management variables include tillage systems, irrigation, and rotations. Once the crop response and the important variables are identified, then the recommendation model should be tested over a range of environmental conditions and N components.

For making fertilizer N recommendations, at least two strategies emerge, depending on other characteristics of the production system. The main distinction between these strategies is the ability of the producer to apply N fertilizer easily at any time during the growing season versus situations that essentially require all the fertilizer to be applied while the plants are small

Table 3. An example data set from Nebraska to determine N fertilizer recommendations for corn with two irrigation systems

| Data | Conventional | Center-Pivot |
|--|--------------|---------------|
| Residual soil N | | 4004 45400 AC |
| 0-20 cm (mg kg ⁻¹ nitrate-N) | 13.5 | 12.3 |
| 20-90 cm (mg kg ⁻¹ nitrate-N) | 7.2 | 4.5 |
| 0-90 cm (wt. average, mg kg ⁻¹ nitrate-N) | 8.8 | 6.2 |
| Organic matter | | |
| 0-20 cm (%) | 1.77 | 1.9 |
| Irrigation water (mg l ⁻¹ nitrate-N) | 30 | 30 |
| Expected yield (mg ha ⁻¹) | 12.6 | 12.6 |
| Estimated water application (cm) | 25 | 25 |
| Estimated crop N need (kg ha ⁻¹) | 309 | 309 |
| 21.4 (expected yield) + 39 | | |
| N credits | | |
| Residual soil N (kg ha ⁻¹) | 79 | 56 |
| 9 (mg nitrate-N kg ⁻¹) | | |
| Organic matter (kg ha ⁻¹) | 56 | 60 |
| 2.5 (expected yield)(% organic matter) | | |
| Irrigation water (kg ha ⁻¹) | 75 | 75 |
| 0.1 (cm depth)(mg l^{-1} nitrate-N) | | |
| Others | | |
| Soybeans (50 kg ha ⁻¹) | 0 | (|
| Alfalfa (135–170 kg ha^{-1}) | 0 | į |
| Manures (variable | 0 | |
| Fertilizer N required (kg ha ⁻¹) | 99 | 111 |

to avoid mechanical damage. Situations that are limited to preplant or sidedress N application involve a greater degree of anticipation on the part of the farmer, because they provide little opportunity to compensate for atypical climatic conditions. In contrast, situations where fertigation or other means of N application (i.e. high clearance vehicle, foliar application, etc.) are possible throughout the growing season allow producers to move closer to spoon-feeding their crops. As it turns out, the greater flexibility a producer has in terms of N application times and forms, the more decisions that must be made about fertilizer costs, labor, and convenience. The two scenarios for N recommendations that follow are for irrigated corn in Nebraska (Table 3). One field is under conventional furrow irrigation and the other adjacent field has been under center-pivot irrigation for several years. Because of the methods of irrigation and past fertilizer N management practices, the levels of residual soil N are different for the two fields.

Conventional field

Consider applying approximately 20–30 kg N ha⁻¹ as a starter fertilizer at planting and the remainder as preplant or sidedress application. Local regulations require preplant applications in excess of 90 kg N ha⁻¹ use a nitrification inhibitor. The credit given for the amount of N supplied in irrigation water is conservative, because the amount of water applied by furrow irrigation can be quite variable within a field, ranging from 60–90 cm. However, because a considerable portion of the irrigation water is applied to corn after silking, a portion of the N in the irrigation water will not be utilized by the crop. The value of 25 cm water application is based on the evapotranspiration minus long-term average precipitation during the growing season.

Center-pivot field

Consider applying approximately 20–30 kg N ha⁻¹ as a starter fertilizer at planting and 50–75% of the remainder as preplant or sidedress application. Use chlorophyll meters or other means to monitor crop N status during the growing season. If an N deficiency appears to be developing, apply 20–30 kg N ha⁻¹ as urea ammonium nitrate in the irrigation water as needed. No N should be applied in the water beyond 3 weeks after silking.

One future N management scenario under development involves the use of equipment to apply variable rates of N. This approach will depend on a field location system, typically a global positioning system (GPS), to identify the position of the fertilizer applicator in the field. The variable-rate map of fertilizer N recommendations will be generated based on a variety of factors, such as residual soil N, organic matter content, and yield goal, for each part of the field. This is an emerging technology that offers the potential to reduce N applications in fertile areas of the field and maintain productivity in less fertile areas.

Conclusions

Our ability to make fertilizer recommendations is becoming more sophisticated, as we learn more about the soil/plant/water system. In the past, yield was the only measure of crop N need; however, this has contributed to some environmental problems. We now know that accounting for existing and projected amounts of available N in the root zone can improve the accuracy of N recommendations. Biologically fixed N, from associative and free-living systems, needs to be quantified, especially in tropical systems. In most cases, maximum yield is not desirable, because it results in economic and environmental risks. Realistic yield expectations should be stressed when making N fertilizer recommendations. We should realize that we cannot achieve or expect 100% efficiency from N fertilizers. Some losses through denitrification and leaching must be expected, and, in some years, unusual weather events will create situations that can significantly reduce or increase N use efficiency.

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